



OPEN

DATA DESCRIPTOR

Plot and landscape-level estimates of tree biomass and carbon stocks in Panama's mangrove Important Bird Areas

Jorge Hoyos-Santillan^{1,2,3,4}✉, Alejandro Miranda⁵, Juliana Chavarría^{2,6}, Carlos Hormazabal⁵, Blas Mola-Yudego⁴ & Esperanza González-Mahecha⁷

We present raw tree inventory data from 41 permanent plots in mangrove forests located in two Important Bird Areas along Panama's Pacific coast: the Bay of Parita and the Bay of Panama. The dataset provides measurements for estimating aboveground biomass and carbon stocks in living trees, following standard inventory guidelines. In addition, we present continuous maps for aboveground biomass and carbon density for the study areas. These data help address uncertainties in Panama's mangrove carbon stock estimates, currently ranging from 25 to 71 MtC, by providing ground-based measurements and landscape-level estimates. The measurements also enable validation of landscape-scale remote sensing approaches, such as NASA's Global Ecosystem Dynamics Investigation (GEDI), essential for regional carbon mapping. As Panama's primary carbon sink is the Land Use, Land-Use Change, and Forestry sector, and the country has included mangrove restoration in its Nationally Determined Contributions, this dataset contributes to national climate strategies. Furthermore, it can serve as a foundation for developing a regional carbon monitoring database for Central American and Caribbean countries currently conducting mangrove inventories.

Background & Summary

Nature-based solutions (NbS) have emerged as a key component of global climate strategies, offering cost-effective approaches to address both climate and biodiversity crises^{1,2}. Among NbS, mangroves are particularly valuable ecosystems whose conservation, restoration, and management can contribute significantly to climate change mitigation and adaptation in the tropics³. These ecosystems are essential for preserving terrestrial and marine biodiversity while protecting coastlines⁴.

Globally, mangroves cover 14.7 Mha⁵, storing between 5.2 and 8.6 gigatons of carbon (GtC)⁶. In addition, mangroves are among the ecosystems with the highest carbon sink capacity³, sequestering between 9.6 to 15.8 MtC annually⁶. Despite their relevance as carbon reservoirs-sinks and biodiversity regulators, mangroves are threatened by land use change (e.g., urban, agriculture, aquaculture) and climate change (e.g., severe droughts, sea level rise)⁷. Consequently, mangrove restoration has gained prominence as a climate change mitigation strategy⁸.

In Panama, where forests cover more than 60% of the territory⁹ and the Land Use, Land-Use Change, and Forestry (LULUCF) sector serves as the sole net carbon sink maintaining the country's carbon-negative status⁹, NbS are vital for the climate strategy. Furthermore, through millions of years, Panama has been the door, bridge, and home for thousands of species crossing between North and South America¹⁰. Panama, as biodiversity bridge, is an structural component of the intercontinental flyway pathways, through which millions of birds migrate every year^{11–13}. Recognizing this importance, Panama has incorporated mangrove restoration into its Nationally Determined Contributions¹⁴.

¹Smithsonian Tropical Research Institute, Panama City, Panama. ²Audubon Americas, New York, USA. ³School of Biosciences, University of Nottingham, Sutton Bonington, UK. ⁴School of Forest Sciences, University of Eastern Finland, Joensuu, Finland. ⁵Departamento de Ciencias Forestales, Universidad de La Frontera, Temuco, Chile. ⁶United Nations Development Programme, Costa del Este, Panama City, Panama. ⁷Inter-American Development Bank, Panama City, Panama. ✉e-mail: jorge.hoyos@umag.cl

Panamanian mangroves cover 0.187 Mha of the country; the majority is located in the Pacific, while only 3% occurs in the Caribbean¹⁵. Current total carbon stock estimates for these ecosystems in Panama vary widely, ranging from 25 to 71 MtC^{15–18}. This uncertainty arises from multiple factors, including highly variable mangrove cover estimates; limited availability of ground data (species composition, tree height and diameter, wood specific gravity and carbon content); the inclusion of different carbon pools (living trees, dead trees, wood debris, root biomass, and soil carbon considering distinct depths (0.3 to 1 m); and methodological variations in calculating carbon stock densities (tC ha⁻¹). Specifically for living tree carbon stocks, variations in sampling methodology (site selection and stratification), experimental plot design (size, shape, expansion factors), and the application of different allometric models (general or site/species-specific) significantly affect carbon stock estimates, complicating cross-study comparisons. In addition, ground data are essential for calibrating and validating remote sensing methodologies for landscape-level biomass and carbon estimates (e.g., NASA's Global Ecosystem Dynamics Investigation (GEDI)) that can be used to develop more ambitious national conservation/restoration programs or develop science-based quantitative pledges for Nationally Determined Contributions. Thus, access to raw data from mangrove forest inventories combined with landscape-level continuous maps could lead to the development of a national open-access dataset allowing the standardization of the criteria for mangrove aboveground biomass estimations and regional data comparison. However, mangrove inventories present unique challenges, requiring careful timing around tidal cycles and substantial resources (e.g., human and financial).

This manuscript presents raw data for estimating aboveground biomass and carbon in living trees from two major Important Bird Areas (IBAs) in Panama's Pacific region: the Bay of Parita and the Bay of Panama. The dataset encompasses measurements from 41 permanent plots established following standard guidelines as part of a regional carbon baseline assessment^{19–21}. In addition, we present continuous maps for aboveground biomass and carbon density for the study areas. Together, these datasets can serve as a foundation for broader regional datasets, supporting ongoing national mangrove inventories across Central America and the Caribbean (e.g., the Regional Blue Carbon Monitoring, Reporting, and Verification Mechanism (MRV) funded by the UK-Blue Carbon Fund in Jamaica, Trinidad and Tobago, Suriname, Colombia, and Panama).

Methods

This study utilized ground data collected during the “Valuing, Protecting and Enhancing Coastal Natural Capital in Panama (PN-T1233) / Blue Natural Heritage” project, implemented by Audubon Americas and funded by the UK Blue Carbon Fund through the Department for Environment, Food and Rural Affairs (DEFRA) via the Inter-American Development Bank (IDB). Field data collection was conducted between 2021 and 2024, following standardized sampling protocols for plot location, design, and measurements^{19–21}. Detailed field and laboratory protocols adhered to established mangrove measurement guidelines^{20,22}.

We collected data on tree diameter, height, and wood specific gravity (WSG) to estimate plot-level aboveground biomass and carbon density in living trees. This methods section describes: i) the sampling design and laboratory analyses used to generate raw data for calculating plot-level aboveground biomass density (AGBd) and aboveground carbon density (AGCd), and ii) the methodology for estimating and validating landscape-level AGBd and AGCd using remote sensing techniques.

In situ sampling design. To allocate plots, we first defined the study areas based on existing shape boundaries for Important Bird and Biodiversity Areas in Panama; specifically, those located in the Bay of Parita and the Bay of Panama (Fig. 1). Next, we stratified the study areas considering two mangrove geomorphic typologies: marine (open coast) and riparian; marine mangroves were situated in open-coast, marine-edge settings, whereas riparian mangroves were located in the margins of rivers and streams^{23,24} (Fig. 1). Then, we opted for a random plot placement using a 500 by 500 m grid placed on top of the mangrove shape developed by the Ministry of the Environment, including a 180 m fringe-buffer area. Finally, we applied a criteria of a 1 km minimum distance among plots. We established 41 plots in total (Fig. 1, Table S1).

Each plot consisted of a linear transect consisting of six circular subplots distributed perpendicular to the sea or river shore. Each subplot had a 7 m radius and were separated by a 25 m distance between the center of each subplot (Fig. 2)^{19,21}. Within each subplot, a nested 5 m radius plot was also established.

Mangroves aboveground biomass density (AGBd)

Vegetation survey. Within each 7 m radius subplot, we identified, measured, and tagged all trees with stems >5 cm diameter at breast height (DBH at 1.3 m height). Specifically, we recorded mangrove species, measured DBH in cm using diameter tapes (Forestry Suppliers, metric fabric-283D/10 M), and measured total height (H) in m using analogue clinometers (Suunto, PM5/66PC). Within the 5 m radius subplots, we measured the same parameters for trees <5 cm DBH. Trees >50 cm DBH were measured within 40 × 125 m (0.5 ha) rectangular subplots positioned along the main transect and encompassing the 7 m radius circular subplots.

Wood specific gravity. We collected wood samples from living trees using an increment borer (3-thread, 5.15 mm, 0.4 m; Jim-Gem, Jackson, MS, US) to estimate wood density (i.e., wood specific gravity, WSG). Fresh volume was measured by water displacement using 5 cm wood segments and wood dry weight was determined after drying at 100 °C (48 h) to calculate WSG₁₀₀^{25,26}. WSG was calculated as the ratio between wood dry weight and fresh volume (g cm⁻³). Using all the segments from each core, WSG was weighted by cross sectional areas to take into account the radial increase of wood volume from the tree center²⁶.

Plot-level forest biomass and carbon density. *Living biomass.* We estimated total aboveground biomass of living trees (AGB_{LT}) using the general equation developed for moist mangrove forest stands²⁷. The allometric equation is: $AGB = 0.0509 \cdot (\rho \cdot DBH^2 \cdot H)$, where AGB_{LT} = Aboveground biomass of living trees (kg),

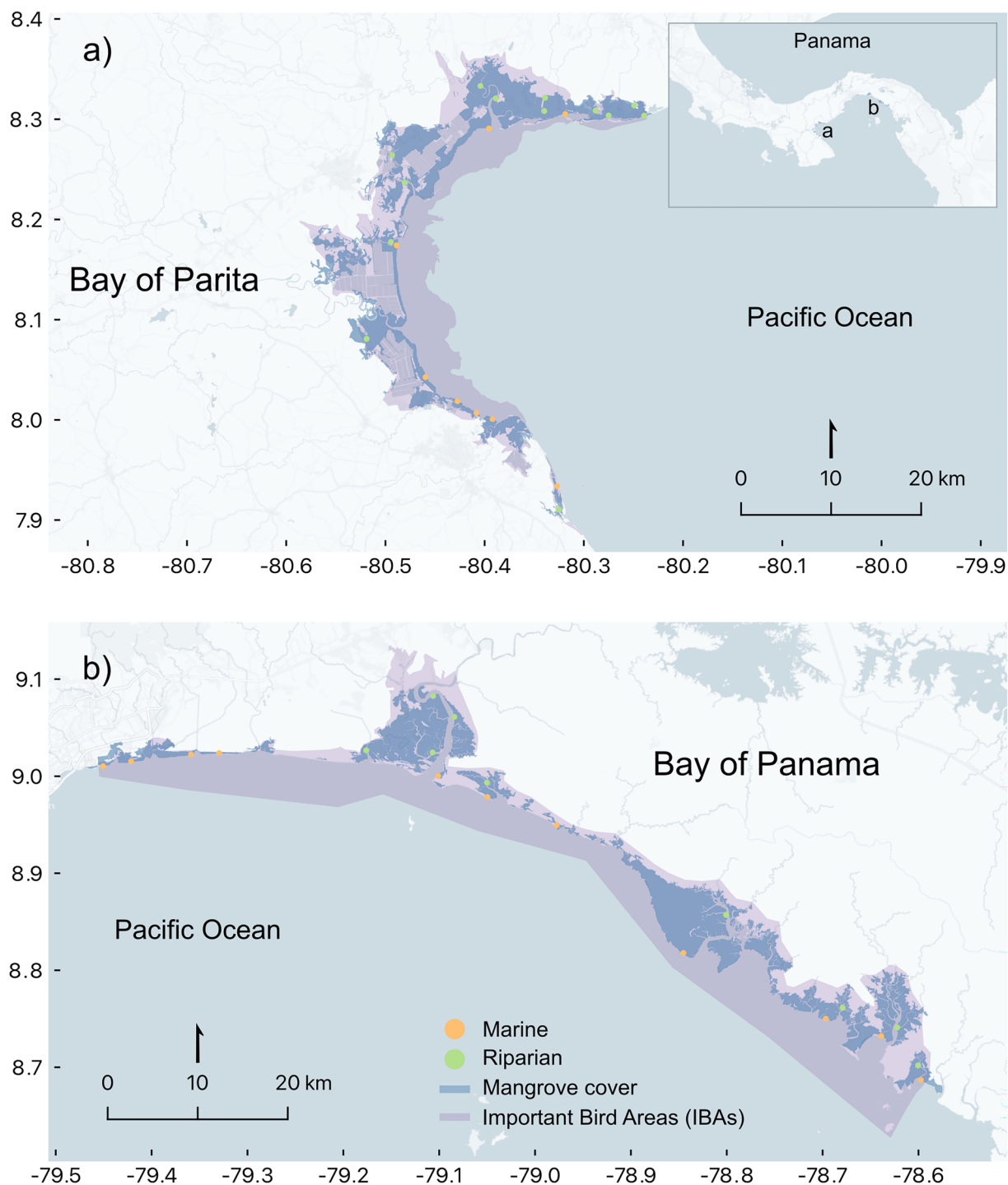


Fig. 1 Plot locations in the study areas: **(a)** Bay of Parita and **(b)** Bay of Panama. Marine plots (orange) and riparian plots (green) are shown overlaid on mangrove cover (blue) and Important Bird Areas (IBAs, purple).

ρ = WSG₁₀₀ (g cm⁻³), DBH = diameter at breast height (cm), and H = total height (m). For ρ , we used the average measured WSG₁₀₀ value for each of the five species for which we collected wood cores (i.e., *Rhizophora mangle*, *Rhizophora racemosa*, *Laguncularia racemosa*, *Avicennia germinans*, and *Avicennia bicolor*). For unmeasured species, we used data available in the literature: for *Mora oleifera*, we used WSG data from the global wood density database^{28,29} and for *Pelliciera rhizophorae*, we used data from Panama's Pacific region³⁰ (Table S2). Using species specific WSG values influences the AGBd and the overall carbon stock estimates, as WSG varies significantly among mangrove species (Table S2). This impact is particularly relevant in areas inhabited by heterogeneous mangrove communities composed of species with contrasting WSG (e.g., *R. racemosa* with WSG $\approx$ 0.91 g cm⁻³ and *P. rhizophorae* with WSG $\approx$ 0.45 g cm⁻³).

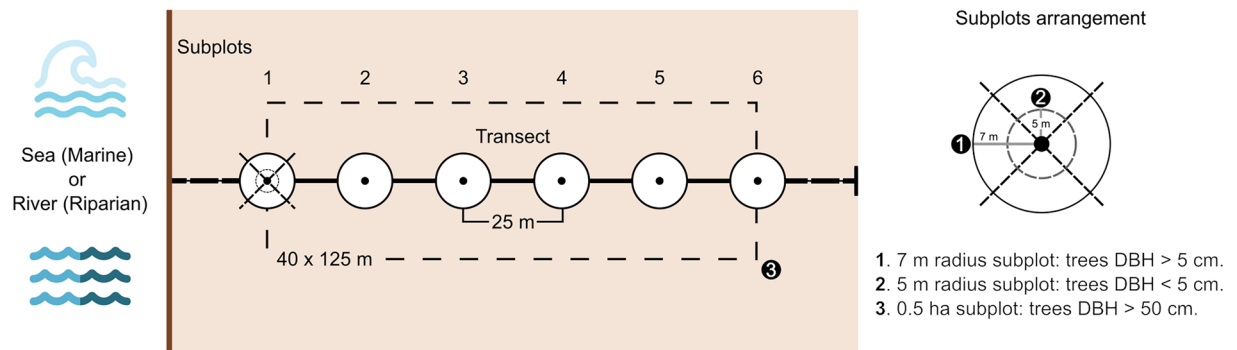


Fig. 2 Schematic plot layout for aboveground biomass estimation in Panamanian mangroves, modified from Kauffman and Donato (2012).

AGB_{LT} (t, tonnes) was calculated for each tree individually. Then, AGB_{LT} of all trees within all circular (i.e., 6 subplots with 7 m radius and 6 nested subplots with 5 m radius plots) and rectangular subplots were summed, independently, for the total area of each subplot type to obtain the total AGB_{LT} per subplot type. The AGB_{LT} values per subplot type were then multiplied by the corresponding expansion factor to convert AGB_{LT} to Aboveground Biomass Density (AGB_d , $t\ ha^{-1}$). The expansion factor for each subplot type was calculated as reciprocal of the total area (ha) of each subplot type (i.e., 6 circular plots of 7 m radius = $923\ m^2 \approx 0.092\ ha$; expansion factor = 10.82; 6 circular plots of 5 m radius = $471\ m^2 \approx 0.047\ ha$; expansion factor = 21.22; and 1 rectangular plot of $40 \times 125\ m = 5000\ m^2 \approx 0.5\ ha$, expansion factor = 2). Finally, the independent AGB_d values per subplot type were added to obtain the overall plot-level AGB_d .

Forest biomass carbon density. We determined biomass carbon content by using a biomass-carbon conversion factor of $0.5\ gC\ g_{biomass}^{-1}$ following standard protocols^{19,20}. This factor was used to convert aboveground biomass density ($t_{biomass}\ ha^{-1}$) (AGB_d) into aboveground carbon density ($tC\ ha^{-1}$) (AGC_d).

Remote sensing. Landscape aboveground biomass density data. We used Global Ecosystem Dynamics Investigation (GEDI) data for landscape-scale forest biomass density analysis (<https://gedi.umd.edu/>). GEDI's satellite Light Detection and Ranging (LiDAR system), operational since late 2018, provides Level-4A products for footprint-level aboveground biomass estimation at 25 m resolution³¹. GEDI footprints are not spatially continuous and are spaced 60 m along-track and 600 m across-track producing a well-defined satellite-based gridded and geolocated spatial sampling pattern for each visit date³². The GEDI L4A product is derived from correlations between AGB ground data, Airborne Laser Scanning (ALS) data, and GEDI footprints developed in study areas throughout the world. First, to predict AGB using GEDI satellite data, ALS data is transformed into the relative heights (RH) measured by GEDI. Subsequently, a model relating plot-level aboveground biomass estimates and GEDI RH metrics is constructed using machine learning algorithms. This model is then applied to predict biomass for each GEDI footprint. The GEDI biomass database adopts the framework established by the Australian Terrestrial Ecosystem Research Network (TERN) Biomass Plot Library, in which above-ground biomass includes only living trees³³.

Landscape-level aboveground biomass density modeling. Since GEDI footprints are not spatially continuous, it was necessary to develop a continuous landscape-level model for AGB_d in the study areas. However, GEDI's data require pre-processing before they can be used to develop a spatially continuous model. GEDI's dataset includes quality bands that assess waveform reliability for capturing three-dimensional attributes of general GEDI products. We applied the following quality filters: `degrade_flag`, `quality_flag`, `run_flag`, `sun_elevation` < 0, and GEDI's L2A and L2B power beams filters 5, 6, 8, and 11, which are recommended for dense vegetation³¹. We selected data with `degrade_flag` = 0, indicating undegraded pointing and positioning information, alongside `quality_flag` = 1, denoting valid waveforms³⁴. Additionally, we utilized `run_flag` = 1, which indicates good waveform fidelity for AGB_d estimation³¹. We excluded GEDI data where the absolute difference between the lowest ground return from the GEDI footprint and NASA's Digital Elevation Model (DEM) at 90 m spatial resolution from the Shuttle Radar Topography Mission (SRTM) exceeded 75 m³⁵.

The landscape-level biomass density model utilized filtered GEDI Level-4A footprints as response variables and multi-source remote sensing data as predictors³⁶. For predictors, we incorporated 22 variables derived from: Sentinel-1 C bands and Synthetic Aperture Radar (SAR) vegetation indices (HH band, HV band, RFDI, and RFDI SD in 100 m radius around each pixel); Sentinel-2 bands and spectral indices (B2, B3, B4, B5, B6, B7, B8, B8A, B11, B12, NDVI, NDVI SD in 100 m radius around each pixel, SAVI, NBR, and EVI); topographical descriptors of the mangroves (elevation³⁷, slope³⁷, and distances to sea and rivers³⁸); and global canopy height³⁹ (Table S3). We constructed a Random Forest machine learning model using 70% of the data for training and 30% for testing. Model performance was evaluated using R^2 and root mean squared error (RMSE) (Fig. 6). All analyses were conducted using Google Earth Engine⁴⁰.

Data Records

The datasets, GeoTIFF files, and code are available at Figshare repository SDATA-25-00610⁴¹.

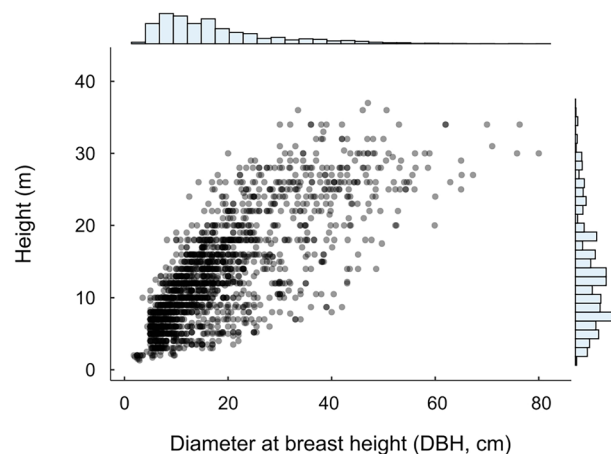


Fig. 3 Relationship between diameter at breast height (DBH, cm) and total height (m) for all mangrove species across the Bay of Parita and Bay of Panama study areas.

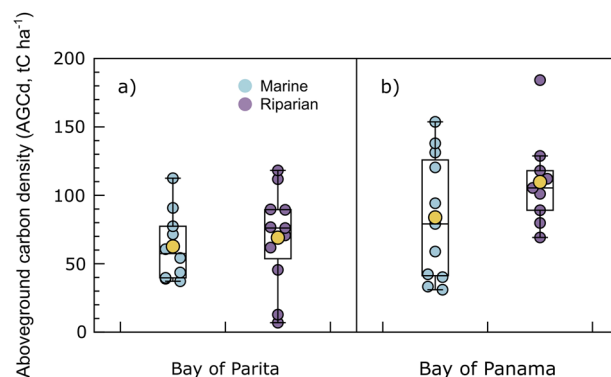


Fig. 4 Above ground carbon density (AGCd, tC ha^{-1}) for mangrove typologies (blue = marine and purple = riparian): (a) Bay of Parita and (b) Bay of Panama. Mean is presented as yellow symbols.

The data records include: (i) the raw data to calculate the plot-level forest aboveground biomass and carbon density, (ii) the aboveground biomass and carbon density estimates for the 41 plots, (iii) the aboveground biomass and carbon density estimates at landscape-level, and (iv) the continuous aboveground biomass and carbon density maps. Specifically, datasets include: (i) CSV (comma separated values) file with columns representing: Registry number = continuous number for each measured tree; Study area = Bay of Parita (ParB) or Bay of Panama (PTYB); Latitude; Longitude; Typology = Marine or Riparian; C pool = carbon pool (aboveground); Sub-C pool = Living biomass; Plot = each of the 41 plots; Subplot = 1 to 6 on each plot; plotId = Plot + Subplot; Diameter class = > 5 cm DBH, < 5 cm DBH, > 50 cm DBH; Family = mangrove classification; Genus = mangrove classification; Species = mangrove classification; SpeciesAbb = abbreviation for each species; DBH (cm) = Tree Diameter at Breast Height measured *in situ*, expressed in cm; Height (m) = Tree total height measured *in situ* with clinometer, expressed in meters (Fig. 3). This raw data, in conjunction with the wood specific gravity (*i.e.*, wood density) (Table S2), allows calculating AGBd ($\text{t}_{\text{biomass}} \text{ha}^{-1}$) and AGCd (tC ha^{-1}) at subplot or plot-level; (ii) CSV file with columns presenting: Study area = Bay of Parita (ParB) or Bay of Panama (PTYB); Plot = each of the 41 plots; Latitude; Longitude; Typology = Marine or Riparian; AGBd ($\text{t}_{\text{biomass}} \text{ha}^{-1}$) and AGCd (tC ha^{-1}) at plot-level (Fig. 4); (iii) CSV file with columns presenting: Study area = Bay of Parita (ParB) or Bay of Panama (PTYB); Latitude; Longitude; AGBd ($\text{t}_{\text{biomass}} \text{ha}^{-1}$) and AGCd (tC ha^{-1}) at landscape-level from GEDI data; (iv) GeoTIFF files (TIFF) with AGBd ($\text{t}_{\text{biomass}} \text{ha}^{-1}$) and AGCd (tC ha^{-1}) at landscape-level for the Bay of Parita and the Bay of Panama (Fig. 5), and (v) ZIP files containing the code to develop the landscape-level AGBd and AGCd estimations for the Bay of Parita and the Bay of Panama (Fig. 5).

Technical Validation

DBH and H measurements are subjected to observational and systematic errors common in forest inventories^{42,43}. For mangrove H records, previous studies have estimated a 7.7% error when using trigonometric equipment such as the clinometers used for this study⁴⁴. The technical validation of our data descriptor focused on examining diameter-height relationships to identify potential measurement biases or errors (Fig. 3). We analyzed the slenderness ratio (total height/DBH, expressed in meters) of all measurements. Of the total

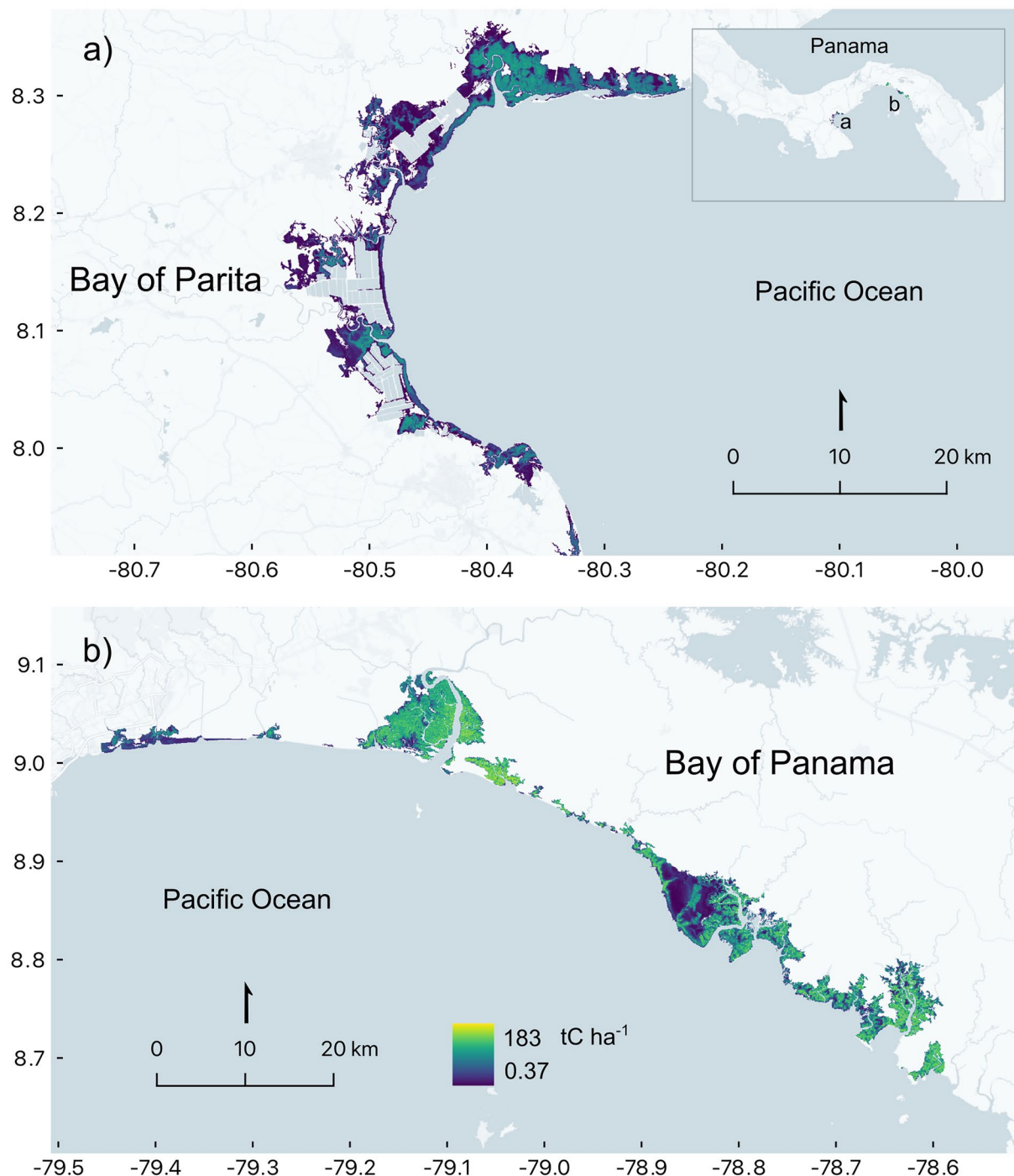


Fig. 5 Above ground carbon density at landscape-level (AGCd, tC ha⁻¹) for (a) the Bay of Parita and (b) Bay of Panama study areas.

measurements, 6% showed extreme values: 39 measurements exhibited excessive slenderness (>150) and 75 showed very low slenderness (<30).

To assess landscape-level biomass estimations performance, we trained our Random Forest model using 70% of our data and evaluated it with the remaining 30%, obtaining an $R^2 = 0.78$ and a Root Mean Square Error (RMSE) of 49 t_{biomass} ha⁻¹ (Fig. 6a). Next, we compared our GEDI-derived AGBd model to the plot-level forest AGBd ground measurements. The relationship between modeled and ground measurements across both study areas (Bay of Parita and Bay of Panama) yielded an R^2 of 0.57 and a (RMSE) of 58 t_{biomass} ha⁻¹ (Fig. 6b). We then computed the global Moran's I on the residuals to test for spatial autocorrelation for the GEDI-derived AGBd residuals ($N = 1\,476$), $I = 0.05$ ($P < 0.001$, permutation test) and for the plot-level AGBd residuals ($N = 41$), $I = 0.11$ ($P = 0.08$). A spatial correlogram of lag-class Moran's I showed a single peak at the maximum lag ($I = 0.33$, $P = 0.156$). Overall, the analysis indicated no robust long-range autocorrelation. Finally, to mitigate

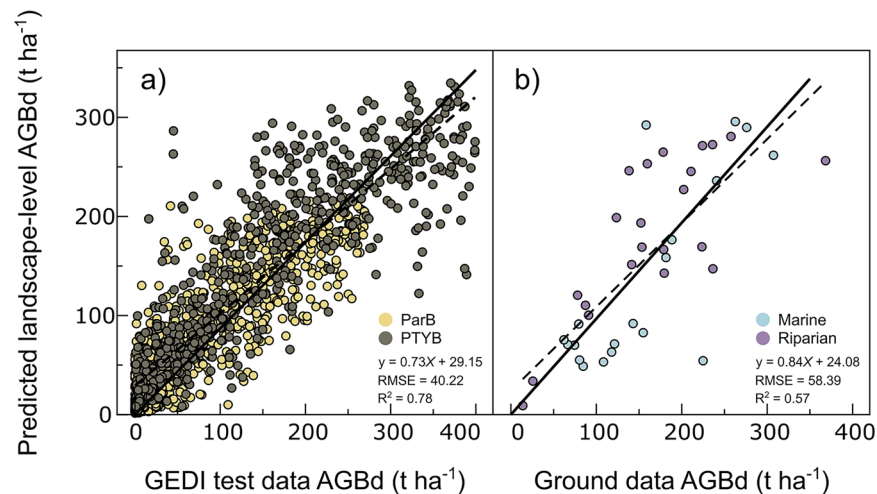


Fig. 6 (a) Random Forest internal evaluation (Symbols represent study areas: yellow = Bay of Parita, dark grey = Bay of Panama; linear regression (dashed line): $R^2 = 0.78$; RMSE = 40.22) and (b) ground/field Aboveground Biomass density (AGBd) data vs predicted landscape-level AGBd ($t_{\text{biomass}} \text{ ha}^{-1}$) (Symbols represent mangrove typology: blue = marine, purple = riparian; linear regression (dashed line): $R^2 = 0.57$; Root Mean Square Error (RMSE) = 58.39 $t_{\text{biomass}} \text{ ha}^{-1}$). The solid lines represent the 1:1 relationship between the variables.

potential geo-location mismatch between GEDI footprints and GPS-recorded plot locations, before comparison, we smoothed both data sets with a 3×3 mean moving-window (*i.e.*, the focal pixel plus its eight neighbors).

Usage Notes

The aboveground biomass density (AGBd, $t_{\text{biomass}} \text{ ha}^{-1}$) and aboveground carbon density (AGCd, tC ha^{-1}) estimates for living trees at plot-level were calculated using the allometric equation: $\text{AGB} = 0.0509 (\rho \text{ DBH}^2 \text{ H})$. This allometric model is adequate for both moist mangrove forest stands and moist forest stands when height (H) is available²⁷. The model was developed including mangrove data for trees with a maximum DBH of 42 cm, whereas the maximum DBH for moist forest trees was 156 cm²⁷. Thus, we applied this model to our entire dataset, as scaling rules for mangroves have been found to be consistent with those for upland moist forest^{45,46}. A carbon conversion factor of $0.5 \text{ gC g}_{\text{biomass}}^{-1}$ was applied (see Methods section). While species-specific allometric models and carbon content values would yield different AGBd and AGCd estimates, the provided raw data enables recalculation using alternative approaches. For the landscape-level AGBd and AGCd estimates, including living trees only, we applied the same carbon conversion factor ($0.5 \text{ gC g}_{\text{biomass}}^{-1}$).

Code availability

R Code and files necessary to develop the landscape-level AGBd and AGCd estimations are available as part of the Figshare files⁴¹ (*i.e.*, Bay of Parita AGBd AGCd landscape model.zip; Bay of Panama AGBd AGCd landscape model.zip).

Received: 5 February 2025; Accepted: 5 June 2025;

Published online: 17 June 2025

References

1. Griscom, B. W. *et al.* Natural climate solutions. *Proc. Natl. Acad. Sci.* **114**, 11645–11650, <https://doi.org/10.1073/pnas.1710465114> (2017).
2. Seddon, N. *et al.* Global recognition of the importance of nature-based solutions to the impacts of climate change. *Glob. Sustain.* **3**, e15, <https://doi.org/10.1017/sus.2020.8> (2020).
3. Taillardat, P., Friess, D. A. & Lupascu, M. Mangrove blue carbon strategies for climate change mitigation are most effective at the national scale. *Biol. Lett.* **14**, 20180251, <https://doi.org/10.1098/rsbl.2018.0251> (2018).
4. Lee, S. Y. *et al.* Ecological role and services of tropical mangrove ecosystems: a reassessment. *Global Ecology and Biogeography* **23**, 726–743, <https://doi.org/10.1111/geb.12155> (2014).
5. Bunting, P. *et al.* Global Mangrove Extent Change 1996–2020: Global Mangrove Watch Version 3.0. *Remote Sens.* **14**, 3657, <https://doi.org/10.3390/rs14153657> (2022).
6. Alongi, D. M. Impacts of Climate Change on Blue Carbon Stocks and Fluxes in Mangrove Forests. *Forests* **13**, 149, <https://doi.org/10.3390/f13020149> (2022).
7. Hagger, V. *et al.* Drivers of global mangrove loss and gain in social-ecological systems. *Nat Commun* **13**, 6373, <https://doi.org/10.1038/s41467-022-33962-x> (2022).
8. Song, S. *et al.* Mangrove reforestation provides greater blue carbon benefit than afforestation for mitigating global climate change. *Nat Commun* **14**, 756, <https://doi.org/10.1038/s41467-023-36477-1> (2023).
9. Ministerio de Ambiente. *Cuarta Comunicación Nacional Sobre Cambio Climático de Panamá*. 273 https://transparencia-climatica.miambiente.gob.pa/wp-content/uploads/2023/08/4CNCC_2023_L.pdf (2023).
10. Hoyos-Santillan, J. *et al.* Panama must protect mangroves and peatlands. *Science* **382**, 654–654, <https://doi.org/10.1126/science.adl3048> (2023).

11. Smith, M. A. *et al.* Bird Migration Explorer. *Natl. Audubon Soc. New York, NY* (2022).
12. Lefebvre, G. & Poulin, B. Seasonal abundance of migrant birds and food resources in panamanian mangrove forests. *Wilson Bull.* **108**, 748–759 (1996).
13. Smith, B. T. & Klicka, J. The profound influence of the Late Pliocene Panamanian uplift on the exchange, diversification, and distribution of New World birds. *Ecography (Cop.)*. <https://doi.org/10.1111/j.1600-0587.2009.06335.x> (2010).
14. Ministerio de Ambiente. *Segunda Contribución Determinada a Nivel Nacional (CDN2)*. 227 <https://unfccc.int/documents/639822> (2024).
15. Ministerio de Ambiente. *Informe Ejecutivo del Mapa de Cobertura Boscosa y Uso de Suelo 2021*. 35 (2022).
16. Bunting, P. *et al.* The Global Mangrove Watch—A New 2010 Global Baseline of Mangrove Extent. *Remote Sens.* **10**, 1669 (2018).
17. Giri, C. *et al.* Status and distribution of mangrove forests of the world using earth observation satellite data. *Glob. Ecol. Biogeogr.* **20**, 154–159, <https://doi.org/10.1111/j.1466-8238.2010.00584.x> (2011).
18. Sanderman, J. *et al.* A global map of mangrove forest soil carbon at 30 m spatial resolution. *Environ. Res. Lett.* **13**, 055002, <https://doi.org/10.1088/1748-9326/aabe1c> (2018).
19. Kauffman, J. B. & Donato, D. C. *Protocols for the Measurement, Monitoring and Reporting of Structure, Biomass and Carbon Stocks in Mangrove Forests*. <https://doi.org/10.17528/cifor/003749> (2012).
20. Howard, J., Hoyt, S., Isensee, K., Pidgeon, E. & Telszewski, M. Coastal Blue Carbon: Methods for Assessing Carbon Stocks and Emissions Factors in Mangroves, Tidal Salt Marshes, and Seagrass Meadows. *Conserv. Int. Intergov. Oceanogr. Comm. UNESCO, Int. Union Conserv. Nature. Arlington, Virginia, USA*. 1–180 (2014).
21. Cifuentes Jara, M. *et al.* *Manual Centroamericano Para La Medición de Carbono Azul En Manglares*. (CATIE, Turrialba, Costa Rica).
22. Kauffman, J. B., Warren, M., Donato, D. C., Murdiyarso, D. & Kurnianto, S. Protocols for the Measurement, Monitoring, & Reporting of Structure, Biomass and Carbon Stocks in Tropical Peat Swamp Forest FIELD HANDBOOK <https://doi.org/10.17528/cifor/003749> (2011).
23. Donato, D. C. *et al.* Mangroves among the most carbon-rich forests in the tropics. *Nat. Geosci.* **4**, 293–297, <https://doi.org/10.1038/ngeo1123> (2011).
24. Worthington, T. A. *et al.* A global biophysical typology of mangroves and its relevance for ecosystem structure and deforestation. *Sci. Rep.* **10**, 14652, <https://doi.org/10.1038/s41598-020-71194-5> (2020).
25. Cornelissen, J. H. C. *et al.* A handbook of protocols for standardised and easy measurement of plant functional traits worldwide. *Australian Journal of Botany* **51**, 335–380 (2003).
26. Smithsonian Tropical Research Institute. ForestGEO Protocols. *ForestGEO* <https://forestgeo.si.edu/protocols> (2015).
27. Chave, J. *et al.* Tree allometry and improved estimation of carbon stocks and balance in tropical forests. *Oecologia* **145**, 87–99, <https://doi.org/10.1007/s00442-005-0100-x> (2005).
28. Chave, J. *et al.* Towards a worldwide wood economics spectrum. *Ecol. Lett.* **12**, 351–366, <https://doi.org/10.1111/j.1461-0248.2009.01285.x> (2009).
29. Zanne, A. E. *et al.* Data from: Towards a worldwide wood economics spectrum. 2123413 bytes [object Object] <https://doi.org/10.5061/DRYAD.234> (2009).
30. Dangremond, E. M. & Feller, I. C. Functional traits and nutrient limitation in the rare mangrove *Pelliciera rhizophorae*. *Aquabot* **116**, 1–7, <https://doi.org/10.1016/j.aquabot.2013.12.007> (2014).
31. Dubayah, R. O. *et al.* Global Ecosystem Dynamics Investigation (GEDI) GEDI L4A Footprint Level Aboveground Biomass Density, Version 2.1. ORNL DAAC <https://doi.org/10.3334/ORNLDAAAC/2056> (2022).
32. Duncanson, L. *et al.* Aboveground biomass density models for NASA's Global Ecosystem Dynamics Investigation (GEDI) lidar mission. *Remote Sensing of Environment* **270**, 112845, <https://doi.org/10.1016/j.rse.2021.112845> (2022).
33. Joint Remote Sensing Research Program. Biomass Plot Library - National collation of stem inventory data and biomass estimation, Australian field sites. Terrestrial Ecosystem Research Network (2021).
34. Mandl, L., Stritih, A., Seidl, R., Ginzler, C. & Senf, C. Spaceborne LiDAR for characterizing forest structure across scales in the European Alps. *Remote Sens. Ecol. Conserv.* **9**, 599–614, <https://doi.org/10.1002/rse2.330> (2023).
35. Fayad, I., Baghdadi, N. & Lahssini, K. An Assessment of the GEDI Lasers' Capabilities in Detecting Canopy Tops and Their Penetration in a Densely Vegetated, Tropical Area. *Remote Sensing* **14**, 2969, <https://doi.org/10.3390/rs14132969> (2022).
36. Bruening, J. M. *et al.* Challenges to aboveground biomass prediction from waveform lidar. *Environ. Res. Lett.* **16**, 125013, <https://doi.org/10.1088/1748-9326/ac3cec> (2021).
37. Copernicus DEM - Global and European Digital Elevation Model. <https://doi.org/10.5270/ESA-c5d3d65>.
38. Sistema Nacional de Información Ambiental. Diagnóstico de Bosques y Otras Tierras Boscosas: año 2023. (2023).
39. Lang, N., Jetz, W., Schindler, K. & Wegner, J. D. A high-resolution canopy height model of the Earth. *Nat. Ecol. Evol.* **7**, 1778–1789, <https://doi.org/10.1038/s41559-023-02206-6> (2023).
40. Gorelick, N. *et al.* Google Earth Engine: Planetary-scale geospatial analysis for everyone. *Remote Sens. Environ.* **202**, 18–27, <https://doi.org/10.1016/j.rse.2017.06.031> (2017).
41. Hoyos-Santillan, J. *et al.* Plot and landscape-level estimates of tree biomass and carbon stocks in Panama's mangrove Important Bird Areas. *Figshare* <https://doi.org/10.6084/m9.figshare.28330115> (2025).
42. Roxburgh, S. H. & Paul, K. I. Comprehensive propagation of errors for the prediction of woody biomass. *Methods in Ecology and Evolution* **16**, 197–214, <https://doi.org/10.1111/2041-210X.14471> (2025).
43. Berger, A., Gschwantner, T., McRoberts, R. E. & Schadauer, K. Effects of Measurement Errors on Individual Tree Stem Volume Estimates for the Austrian National Forest Inventory. *Forest Science* **60**, 14–24, <https://doi.org/10.5849/forsci.12-164> (2014).
44. Saliu, I. S. *et al.* An accuracy analysis of mangrove tree height mensuration using forestry techniques, hypsometers and UAVs. *Estuarine, Coastal and Shelf Science* **248**, 106971, <https://doi.org/10.1016/j.ecss.2020.106971> (2021).
45. Rovai, A. S. *et al.* Macroecological patterns of forest structure and allometric scaling in mangrove forests. *Global Ecology and Biogeography* **30**, 1000–1013, <https://doi.org/10.1111/geb.13268> (2021).
46. Price, C. A. *et al.* Global Data Compilation Across Climate Gradients Supports the Use of Common Allometric Equations for Three Transatlantic Mangrove Species. *Ecology and Evolution* **14**, e70577, <https://doi.org/10.1002/ece3.70577> (2024).

Acknowledgements

The authors would like to acknowledge the support from the National Audubon Society, the Inter-American Development Bank (“IDB”), and Smithsonian Tropical Research Institute through the project “Valuing, Protecting and Enhancing Coastal Natural Capital in Panama (PN-T1233) / Blue Natural Heritage”, funded by Department for Environment, Food and Rural Affairs (DEFRA) UK Blue Carbon Fund. In addition, the authors would like to acknowledge the support from Panama’s Ministry of the Environment for providing the research permits ARB-012-2022, ARB-002-2022, and ARB-112-2024. A.M. acknowledges ANID/Fondecyt Initiation/2024-11240356 and ANID/Fondecyt/2023-IT2310109 funding. Authors would like to thank the field support provided by Javier Vargas, Alicia Sanjur, Kevin Madrid, Eric Manzané, Ronal Rodríguez, Karen Domínguez, Evelyn Jaén, Sozimo Villalobos, Johnny Pardo, Generino Batista, Manuel Saez, and Gabriel Jaramillo.

Author contributions

Conceptualization, investigation, fieldwork: J.H.S., J.C., A.M., E.G.M.; data analysis: J.H.S., J.C., A.M., B.M.Y., C.H.; writing, review, and editing: J.H.S., J.C., A.M., B.M.Y., E.G.M.; funding acquisition: J.H.S., A.M.

Competing interests

The authors declare no competing interests.

Additional information

Supplementary information The online version contains supplementary material available at <https://doi.org/10.1038/s41597-025-05354-5>.

Correspondence and requests for materials should be addressed to J.H.-S.

Reprints and permissions information is available at www.nature.com/reprints.

Publisher's note Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons licence, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons licence, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons licence and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this licence, visit <http://creativecommons.org/licenses/by/4.0/>.

© The Author(s) 2025